



Estimating chlorophyll profiles from electronic tags deployed on pelagic animals

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ABSTRACT: Electronic tags deployed on pelagic animals in the ocean are revolutionizing our understanding of how pelagic species interact with the physical environment, by making simultaneous eco-physiological and *in situ* oceanographic measurements at relatively low cost. Pelagic animals with electronic tags are also increasingly being used as ocean sensing platforms. Here, we demonstrate the concept of estimating *in situ* chlorophyll concentration profiles from light level and depth data collected by electronic tags, using a bio-optical model. We assessed this concept by deploying electronic tags at 14 oceanographic stations in the Pacific Ocean, together with a fluorometer and Niskin bottles. In the euphotic zone, tag-estimated chlorophyll concentrations correlated significantly with simultaneous chlorophyll concentration measurements from filtered water samples ($R^2 = 0.41$, $p < 0.0001$) and a profiling fluorometer ($R^2 = 0.29$, $p < 0.0001$). Below the euphotic zone, peaks in the light attenuation profile corresponded to deep scattering layers, with a 17 m rms difference between the deep scattering layer depths estimated by the tags and an Acoustic Doppler Current Profiler. *In situ* chlorophyll profiles were also derived from electronic tags deployed on Pacific bluefin tuna in Baja California waters, and found to be comparable to chlorophyll profiles in the World Ocean Database for the region and season. Although more research is needed to improve this approach, the results from both shipboard and Pacific bluefin tuna experiments indicate that light level attenuation profiles derived from electronic tags can be successfully used to estimate chlorophyll concentration profiles.

KEY WORDS: Light level · Archival tag · Bio-optical model · Chlorophyll concentration estimation · Deep scattering layer · Pacific bluefin tuna · *Thunnus orientalis*

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INTRODUCTION

The distribution of chlorophyll throughout the water column is one of the most important biological parameters in the ocean because it is an indicator of the spatial and temporal variability of primary productivity (Behrenfeld & Falkowski 1997). In recent years, the increasing capabilities of electronic tags coupled with synoptic observations of the ocean surface from satellite-derived remote-sensing data have improved our understanding of the relationship between the spatio-

temporal distribution of pelagic animals and various environmental parameters, including chlorophyll (Block et al. 2002, Costa et al. 2002, Polovina et al. 2004, Block et al. 2005, Weng et al. 2005, Biuw et al. 2007, Teo et al. 2007). For example, surface chlorophyll distributions estimated by satellite-based optical instruments like the Sea-viewing Wide Field-of-view Sensor (SeaWiFS) were integrated with the movement patterns of pelagic animals derived from electronic tag data to help understand this relationship (Polovina et al. 2004, Teo et al. 2007).

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However, these satellite-borne instruments perform poorly in areas with high cloud cover and/or seasonal ice coverage, and are not able to provide subsurface chlorophyll concentration profiles. Subsurface chlorophyll measurements have traditionally been made from research vessels, using profiling fluorometers and water samples collected by Niskin bottles. Alternatively, chlorophyll profiles can be obtained from fluorometers deployed on fixed moorings or autonomous platforms like Argo floats (Roemmich et al. 2004) and autonomous underwater vehicles (AUVs) (Yu et al. 2002). However, it remains relatively difficult and costly to coordinate research vessels and/or AUVs to sample an area at the same time as pelagic animals are observed to use the area, particularly over long durations (Baumgartner et al. 2003).

Since many pelagic animals dive repeatedly through the water column while traveling over long distances and electronic tags can measure oceanographic parameters at high sampling rates, the electronic tags deployed on pelagic animals can also act as low-cost *in situ* ocean sensing platforms. Thus, there is great interest in the biological and oceanographic research communities to maximize the use of oceanographic data collected under research programs utilizing electronic tags on animals (Boehlert et al. 2001, Fedak et al. 2001, Block et al. 2002, McMahon et al. 2005). Oceanographic parameters that can now be sampled by animal-borne electronic tags include temperature and salinity, but adding other parameters like chlorophyll is of great interest (Biuw et al. 2007, Charrassin et al. 2008). Ideally, electronic tags should incorporate fluorometers in order to obtain simultaneous estimates of chlorophyll in the water column as experienced by pelagic animals. However, there are few, if any, currently commercially available fluorometers with the appropriate size, weight and power consumption characteristics for incorporation into current electronic tags. It would especially be difficult to design fluorometers for use with pelagic fishes like tunas and sharks because the fluorometers would need to be orders of magnitude smaller than current instruments.

Here, we demonstrate an approach to estimate *in situ* chlorophyll concentrations from light and depth data collected by currently available geolocating electronic tags. Light and depth data are routinely collected by electronic tags to estimate the location of pelagic animals (Ekstrom 2004, Teo et al. 2004) but were not originally intended to be used to estimate chlorophyll concentrations. However, we can utilize these routinely collected data to estimate subsurface chlorophyll concentrations by first determining the water column's light attenuation profile from the light and depth data. We can then use a bio-optical model to derive chlorophyll estimates from the light attenuation

profiles. This will allow us to obtain subsurface chlorophyll profiles along an animal's path while simultaneously observing part of the animal's behavior and physiology with the electronic tags. It also has the potential to greatly increase the number of *in situ* chlorophyll profiles available to the oceanographic community and hence improve our understanding of subsurface chlorophyll dynamics.

Our objectives in the present study were to assess the ability of commercially available electronic tags (LTD2310 archival tags, Lotek Wireless) to estimate subsurface chlorophyll concentrations in 2 ways: (1) in relatively controlled conditions by deploying the tags on an oceanographic carousel, and (2) in typical tag deployment conditions by deploying the tags on pelagic fish. Testing this approach by deploying tags on pelagic animals is important because changes in the orientation of the light sensor and/or animal during a dive may affect the light attenuation profiles due to changes in the optical geometry (Mobley 1994) and hence affect the accuracy and precision of the chlorophyll concentration estimates. To accomplish the first objective, we deployed the tags on a standard oceanographic carousel at stations in the tropical Pacific Ocean. Tag-derived chlorophyll concentration estimates were then compared with corresponding estimates from *in situ* water samples and a profiling fluorometer. Since light attenuation and acoustic echo intensity both tend to increase with increased biomass in the water column (Roe & Griffiths 1993), we also hypothesized that the light attenuation profiles from the tags may be comparable to the echo intensity profiles from the shipboard Acoustic Doppler Current Profiler (ADCP). Therefore, we also compared the light attenuation profiles from the electronic tags with corresponding echo intensity profiles from the ADCP at the same stations. To accomplish the second objective, we deployed similar archival tags on Pacific bluefin tuna *Thunnus orientalis* in the eastern tropical Pacific Ocean. After recovering the archival tags, tag-derived chlorophyll concentration estimates were compared with records from the World Ocean Database (WOD) for the same season and location.

MATERIALS AND METHODS

Electronic tags. Standard LTD2310 archival tags (16 mm \varnothing , 76 mm length, 45 g in air; Lotek Wireless) measure and record relative light level, pressure, water temperature, and, during the bluefin tuna experiments, also body temperature. Pressure measurements were converted into depth data by assuming 1 decibar was approximately 0.993 m. The tag has a sensor stalk (user-specified length) that contains a Teflon-coated

optical fiber that is doped with fluorescent dye and channels light to the photodiode light sensor in the tag body. The sensor stalk is highly flexible and is coated with a black, opaque layer except for an optical window near the end of the stalk, which allows light to reach the optical fiber. The excitation spectrum of the fluorescent dye is used as an optical filter with an acceptance band centered on 470 nm (Ekstrom 2004). The acceptance band has a nominal bandwidth of 50 nm and a sharp edge at approximately 500 nm. In pure water, blue light at 450 nm has the lowest attenuation coefficient (0.018 m^{-1}) within the visible light spectrum (Morel & Prieur 1977). The acceptance band of 470 nm was chosen during the development of the tag to reduce the attenuation of light with increasing depth and hence improve the ability to detect light at depth (Ekstrom 2004). The tip of the sensor stalk contains a thermistor that measures the water temperature, and the tag body contains a pressure sensor and a body temperature sensor. In a typical deployment in a tuna, the tag body is surgically implanted into the peritoneal cavity of the fish so that the sensor stalk protrudes through the body wall into the water column (Block et al. 1998). Placement of the tag in the peritoneal cavity of an endothermic animal also keeps the tag at a relatively stable and warm temperature, which may help stabilize the electronics.

The LTD2310 sensor stalk can be regarded as a cylindrical light collector; in a typical deployment on a fish, the flexible stalk will change its orientation rapidly with respect to the ambient light field as the fish moves through the water column. Therefore, we considered the tag-measured irradiance, E_{tag} , to approximate the scalar irradiance at 470 nm, E_o . Irradiance values are recorded by the tag as integers on a logarithmic scale that has nominally 32 units change in reading per order of magnitude change in irradiance (Ekstrom 2004). However, subsequent calibrations of the tags used in the present study showed that the logarithmic scale ranged from 36 to 37 units change per order of magnitude change in irradiance. It is important to note that the LTD2310 tag was designed to measure relative light levels and not absolute light levels. The light level measurements from the tags are therefore not calibrated to absolute levels of irradiance (W m^{-2}), and light levels recorded by one tag may be different from another tag. However, this is not critical because our approach is to use the change of relative light level over depth to calculate the rate of light attenuation in the water column (see 'Model' subsection below for more details). For each tag, the logarithmic scale of the light level readings is verified by the manufacturer to be constant over the entire usable irradiance range (~ 10 orders of magnitude) (Ekstrom 2004). The light sensor is also temperature compen-

sated so that the irradiance values above the sensitivity threshold do not change with temperature. However, the sensitivity threshold varies by up to 1.5 orders of magnitude with changes in temperature (Ekstrom 2004). All irradiance values below the sensitivity threshold are flagged by the onboard processor and removed prior to analysis.

Model. We used a modified version of the bio-optical model developed by Morel (1988) to convert the irradiance values measured by the tags, E_{tag} , into chlorophyll concentration estimates, C_{tag} . Morel (1988) showed that the wavelength-specific diffuse attenuation coefficients of downward irradiance, K_d , of surface Case-I waters can be related to the corresponding chlorophyll concentrations, C , by:

$$K_d \approx K_{\text{dw}} + \chi C^e \quad (1)$$

where K_{dw} is the wavelength-specific diffuse attenuation of downward irradiance due to pure seawater, and χ and e are wavelength-specific fitted constants.

The K_d of the water column is defined as:

$$K_d = -\frac{1}{E_d} \frac{dE_d}{dz} \quad (2)$$

where z is the depth of the irradiance measurement and E_d is the downward irradiance. However, since the LTD2310 tag records relative irradiance values as integers on a logarithmic scale that had nominally 32 units change in reading per order of magnitude change in irradiance, we rearranged Eq. (2) and calculated the diffuse attenuation coefficient of tag-measured irradiance, K_{tag} , by:

$$K_{\text{tag},z} = 1 - 10^{\left[\frac{E_{\text{tag},z+1} - E_{\text{tag},z}}{D} \right]} \quad (3)$$

where $K_{\text{tag},z}$ is the K_{tag} at depth z , D is the number of units change in the E_{tag} reading per order magnitude change in irradiance, and $E_{\text{tag},z}$ and $E_{\text{tag},z+1}$ are the interpolated irradiance values at depths z and $z+1$, respectively. The raw E_{tag} values were first interpolated with respect to depth using a loess fit (Cleveland 1993). An example from our shipboard experiments (described in 'Shipboard experiments') is shown in Fig. 1A. For each 1 m, the interpolated $E_{\text{tag},z}$ and $E_{\text{tag},z+1}$ were then used to calculate K_{tag} from Eq. (3) (Fig. 1B). If we obtained a calibrated D value from the manufacturer for the tag, we would use the calibrated value. However, if the calibrated D value was not available for a particular tag, we parameterized D with a value of 37 because the mean D of the calibrated tags was 37 ± 1 . By rearranging Eq. (1) and replacing constants K_{dw} , χ and e with values for 470 nm (Morel 1988), we estimated the chlorophyll concentration C_{tag} from K_{tag} for each 1 m bin. The C_{tag} values from all the depth bins were then collated to provide the tag-

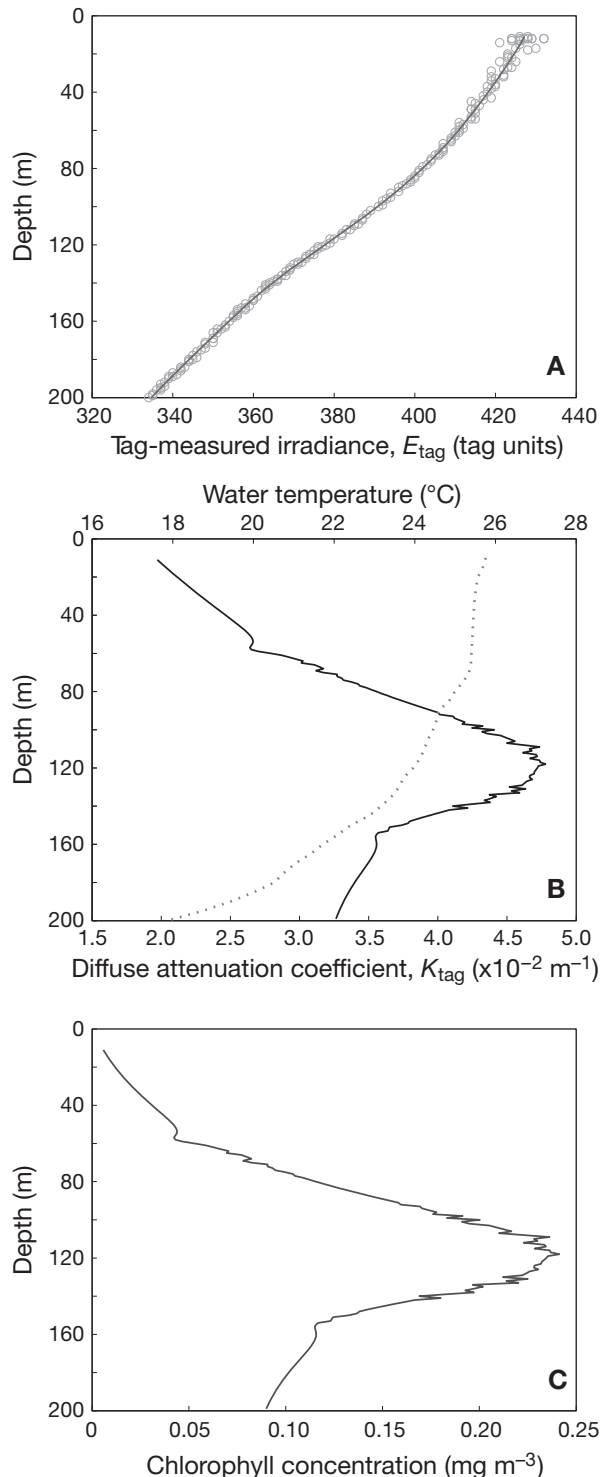


Fig. 1. Irradiance, diffuse attenuation coefficient and chlorophyll profiles from the upper 200 m at Stn S187-045 in 2003, using tag A0586. (A) Irradiance (E_{tag}) measured by tag A0586 (grey circles) and interpolated irradiance profile calculated using a loess fit (solid line). (B) Water temperature profile (dotted line) and the diffuse attenuation coefficient profile of tag-measured irradiance, K_{tag} (black line), which was calculated from the interpolated irradiance profile. (C) Tag-estimated chlorophyll profile, using Eq. (1)

estimated chlorophyll profile (Fig. 1C). This model was coded in Matlab (Matlab R14SP3, Mathworks).

Although the tag measures scalar irradiance rather than downward irradiance, computer simulations show that K_d is well approximated by the diffuse attenuation coefficient of scalar irradiance, K_o ($K_o/K_d = 1.01$ to 1.06), for natural values of scattering and absorbance in the water column (Kirk 1994). Therefore, we expect to overestimate the chlorophyll concentration by up to 6% due to our use of scalar rather than downward irradiance. However, this source of error would not change the shape of the profile substantially because it affects the estimates equally. We consider this level of error to be acceptable because our aim in this study is to demonstrate the feasibility of this approach and approximate the relative chlorophyll profile to the first-order. With further improvements to this approach, we believe that the errors can and will be reduced substantially. We detail the likely sources of error and their effects in the discussion

Shipboard experiments. In 2003 and 2005, we assessed the ability of LTD2310 tags to estimate chlorophyll concentrations during oceanographic cruises on the SSV 'Robert C. Seamans'. Each cruise consisted of a roundtrip between Oahu, Hawaii, and Palmyra Atoll, with 7 oceanographic stations (Fig. 2). At each station, we attached 2 LTD2310 tags (programmed to record every 4 s) to the top of the water-sampling carousel (SBE 32-16, Seabird Electronics), with the sensor stalk protruding into the water column. During a typical deployment on a pelagic fish, the sensor stalk protrudes into the water column from the side of the fish. The orientation of the sensor stalk on the carousel approximated the orientation during a typical deployment on a fish. Our aim for this set of experiments was to assess our approach under relatively controlled conditions with fewer changes in optical orientation than would be expected for the deployments on pelagic fish (see 'Materials and Methods — Pacific bluefin tuna experiments'). The carousel was equipped with 12 2.5 l Niskin bottles (General Oceanics) and an SBE-19plus CTD profiler (Seabird Electronics) with an attached *in situ* chlorophyll fluorometer (SCF, Seapoint Sensors). Oceanographic sampling was conducted at approximately local noon and had maximum depths of 534 ± 50 m (Table 1).

After each cast, the light attenuation (K_{tag}) and chlorophyll concentration (C_{tag}) estimates for each tag were calculated using the method described above, and vertical profiles were constructed from these values. The C_{tag} profiles in the euphotic zone for each station were compared to the chlorophyll profiles from the fluorometer and the water samples taken using the Niskin bottles. We only compared chlorophyll profiles in the euphotic zone because all chlorophyll maxima

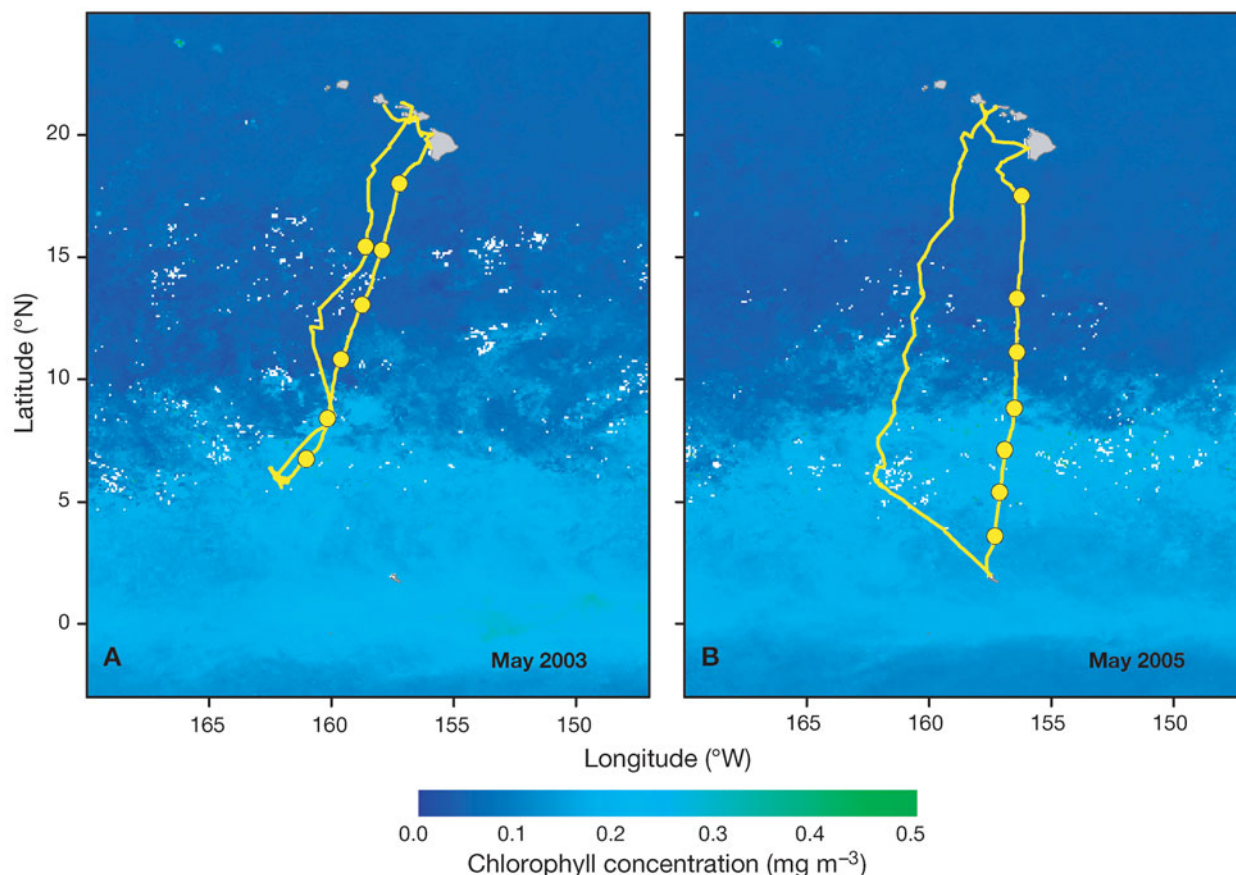


Fig. 2. Cruise track of SSV 'Robert C. Seamans' in (A) May 2003, Cruise S187 and (B) May 2005, Cruise S199. Mean surface chlorophyll concentrations for the corresponding month and year are shown. Circles indicate locations of oceanographic stations; white areas indicate clouds and/or missing data; grey areas indicate land

Table 1. Locations and dates of oceanographic stations during cruises in 2003 and 2005 on the SSV 'Robert C. Seamans'. Start time is in Hawaiian Standard Time

Date	Start time	Stn ID	Longitude (°W)	Latitude (°N)	Depth of cast (m)
2003					
14 May	11:11	S187-014	157.218	18.005	552
15 May	11:17	S187-016	157.920	15.272	434
16 May	11:27	S187-018	158.728	13.045	497
17 May	11:11	S187-021	159.598	10.803	460
18 May	11:31	S187-023	160.155	8.388	637
19 May	11:11	S187-026	161.012	6.737	589
5 Jun	11:23	S187-045	158.600	15.428	530
2005					
14 May	12:06	S199-010	156.080	15.647	521
15 May	12:00	S199-013	156.375	13.328	560
16 May	11:40	S199-016	156.375	11.093	560
17 May	11:56	S199-020	156.513	8.842	517
18 May	12:32	S199-023	156.858	7.103	547
19 May	12:08	S199-026	157.067	5.390	551
20 May	12:00	S199-029	157.298	3.647	556

were within this zone and deep scattering layers affected the quality of irradiance profiles below the euphotic depth. The euphotic zone was defined as the zone between the surface and the euphotic depth, defined here as the penetration depth of 1% of surface irradiance. The mean euphotic depth of the stations was 142 ± 25 m. The irradiance data for the upper 10 m of the water column were not used to estimate K_{tag} and C_{tag} profiles because the bottom of the ship's hull was at approximately 5 m depth and the irradiance data at those depths (<10 m) were relatively noisy due to shading.

Water sample chlorophyll concentrations were determined using standard fluorometric analysis of acetone extracts of the filtered samples. Water samples collected using Niskin bottles were divided into 2 subsamples and filtered (1000 cm^3 in 2003; 250 and 500 cm^3 in 2005) onto glass fiber filters (Whatman GF/F, nominally $0.7 \mu\text{m}$) using positive pressure. The filters were placed in a test tube, wrapped in aluminum foil and frozen in the dark. Chlorophyll was extracted from the filter with 7 ml of HPLC grade acetone for 24 h in the dark. The pigment concentration was then analyzed by the fluorometric method (Yentsch & Menzel

1963) with a blanked and calibrated fluorometer (Turner Designs 10-AU). For each station and depth, we calculated the mean chlorophyll concentration from the 2 subsamples and compared that to corresponding C_{tag} profiles. The C_{tag} profiles were also compared with those made by the *in situ* chlorophyll fluorometer (Astheimer & Haardt 1984). We correlated the C_{tag} profiles with the chlorophyll concentrations from the water samples and the fluorometer and calculated the overall rms difference between the corresponding profiles. The rms differences in the chlorophyll maxima depths detected by the 3 methods were also calculated.

In addition, we used the ADCP (75 kHz Ocean Surveyor, RD Instruments) aboard the vessel to estimate the deep scattering layer depth and compared the ADCP's echo intensity profile with the K_{tag} profile at the oceanographic stations. The acoustic backscatter strength of the water column is related to the biomass profile of the water column and the ADCP's echo intensity profile has been used to estimate the biomass of deep scattering layers (Flagg & Smith 1989, Heywood et al. 1991, Griffiths & Diaz 1996). Since light attenuation also increases with increased biomass in the water column, we hypothesized that the light attenuation profiles from the tags may be comparable to the echo intensity profiles from the shipboard ADCP. We set the ADCP software to aggregate the raw echo intensity of each acoustic beam into ensembles of 20 min and 10 m bins (60 bins, 18 to 608 m). The echo intensity profiles of the 4 beams were then averaged to obtain the echo intensity profile of each ensemble. For each station, we selected the echo intensity profiles from 3 ensembles collected during the cast and compared the mean echo intensity profile with K_{tag} profiles obtained by the tag. The deep scattering layer depth of each station was determined by the depth of the secondary peak of the echo intensity profile. We then calculated the rms difference between the deep scattering layer depths estimated by the ADCP and the tag.

Pacific bluefin tuna experiments. We also assessed this method by deploying 48 LTD2310 archival tags on Pacific bluefin tuna in the eastern tropical Pacific Ocean. The fish were caught by commercial purse seines and transferred by tow pens to grow-out pens on the Pacific coast of Baja California, Mexico. On 9 March 2005, the fish were brought up onto a vinyl surgical pad on the deck of a barge for surgery. Each bluefin tuna was measured (cm curved fork length [CFL]) and an LTD2310 archival tag was surgically implanted into the peritoneal cavity of the fish so that the sensor stalk protruded from the body of the fish (Block et al. 1998). The fish were immediately transferred to a tow pen that was towed offshore (31.718° N, 116.997° W) before the fish were released. All procedures were conducted in accordance with approved

animal care protocols. The tags were programmed to sample every 4 or 8 s. Since the LTD2310 tags were surgically implanted into the peritoneal cavity of the tuna, the fish had to be recaptured to retrieve the tag and the collected data.

Sixteen of the tagged bluefin tuna were subsequently recaptured by commercial fishers. For each day, we extracted the irradiance and pressure data from a period of 2 h surrounding the local noon to minimize variation of the ambient light field (see 'Discussion' for more details). Corresponding C_{tag} values were estimated from the extracted irradiance and depth data using the method described above. Irradiance data from the upper 10 m were not used because the irradiance data at those depths were relatively noisy. Profiles with sparse data (<100 data points) or large spikes (>2.0 mg m⁻³) were eliminated before further analysis. Based on the WOD (www.nodc.noaa.gov) and our oceanographic casts in the area, this value is outside the 95% confidence interval of subsurface chlorophyll maxima concentrations recorded in the area. Individual chlorophyll profiles were smoothed with a robust, locally weighted least squares fit (Cleveland 1993). To visualize how the chlorophyll profiles from a single fish changed over time, we concatenated daily C_{tag} profiles from the same fish into a single array and applied a 2-dimensional 3 × 3 median filter to reduce noise in the array.

Geolocations of the Pacific bluefin tuna were estimated from light level longitude and sea surface temperature (SST) data recorded by the tags using a modification of the methods described by Teo et al. (2004). SST data from the tags were combined with corresponding light level longitude estimates to obtain daily latitude estimates. For a given day, the latitude at which the tag-recorded SSTs best matched the corresponding remotely sensed SSTs, along the light level longitude estimate, was considered the latitude estimate for the day. Based on swimming speeds estimated from acoustic tracking studies (Boustany et al. 2001), the SST matching process for each day was constrained to the area that the fish could have realistically moved. We also improved the geolocation algorithm with the addition of the capacity to switch to cloud-free, microwave-based, remotely sensed SST data (TMI/AMSRE, <ftp://ftp.discover-earth.org/sst/>) during periods with high cloud cover, and an improvement to the algorithm used ensured the movement path did not cross land. Daily maximum diving depths recorded by the tags were also used to filter geolocation estimates so that the maximum diving depth did not exceed the known bathymetry (inclusive of error estimates) at the geolocation estimate for the corresponding day. Archival tags were found to have rms errors of 0.78° and 0.90° for longitude and latitude estimates, respectively (Teo et al. 2004).

Mean depths of the chlorophyll maxima and mean chlorophyll concentrations at the maxima estimated from electronic tags were compared to the profiles recorded in the WOD for the season and region. Here, we used C_{tag} profiles that went deeper than the euphotic depth to ensure that the profiles included the chlorophyll maxima. We extracted corresponding WOD chlorophyll profiles for the same month that were within 1° of the geolocation estimate. However, the WOD profiles were typically from a different year. It was therefore not possible to determine the actual errors in the tag-derived chlorophyll estimates because the chlorophyll data from the WOD were from a different year and therefore not the actual chlorophyll concentration when the bluefin tuna was in the area. However, this was not critical because we did not set out to determine the error distribution of the chlorophyll estimates. Rather, our aim for this set of experiments was to determine if we could derive reasonable chlorophyll profiles for this area and season using these electronic tags. We assumed that the chlorophyll profiles in the WOD for the same season and region provided a reasonable range of chlorophyll maxima concentrations and depth likely to be experienced by the tuna. We subsequently calculated the overall rms differences between the chlorophyll maxima concentrations and depths from the tags and the WOD. In addition, we determined if the chlorophyll maxima concentration and depth of each profile were within the 95 % confidence interval of the corresponding WOD mean chlorophyll maxima concentration and depth.

RESULTS

Shipboard experiments

In the euphotic zone, chlorophyll profiles estimated by the LTD2310 tags were representative of the oceanographic stations and were similar to the profiles measured from the fluorometer and water samples. As an example, the chlorophyll concentrations determined by all 3 methods at Stn S187-045 are shown in Fig. 3. Across all 14 stations, tag-estimated chlorophyll concentrations were significantly correlated with chlorophyll concentrations measured from the fluorometer ($R^2 = 0.29$, $n = 142$, $p < 0.0001$) and water samples ($R^2 = 0.41$, $n = 142$, $p < 0.0001$) (Fig. 4). However, there were clearly differences between the methods. Overall rms differences between the euphotic zone chlorophyll concentrations estimated by the tag versus corresponding chlorophyll concentrations measured from the filtered water samples and fluorometer were 0.088 and 0.27 mg m^{-3} ($n = 142$), respectively. This

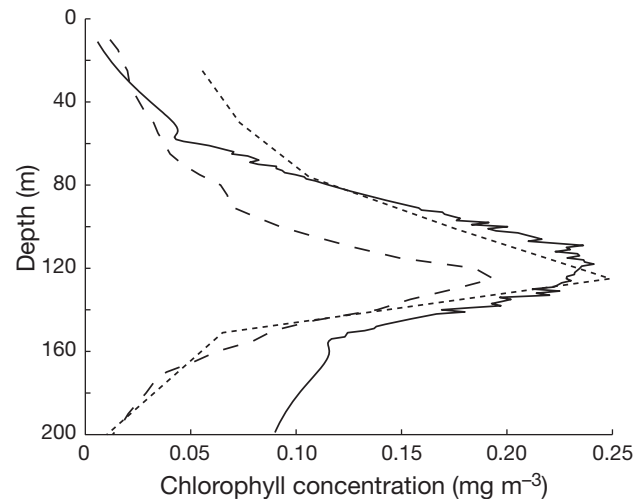


Fig. 3. Chlorophyll profiles from the upper 200 m at Stn S187-045, using tag A0586 (—), filtered water samples (-----) and profiling fluorometer (- - -)

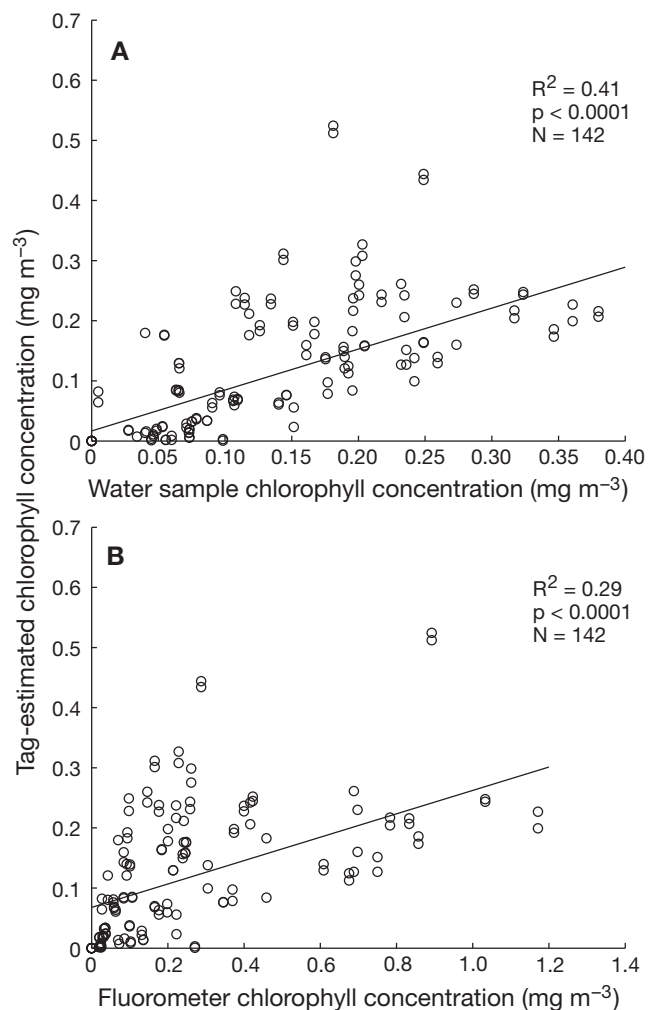


Fig. 4. Correlation of tag-estimated chlorophyll concentrations with (A) filtered water samples and (B) profiling fluorometer from all oceanographic stations

compares well with the rms difference (0.23 mg m^{-3} , $n = 142$) between corresponding chlorophyll concentrations measured from the filtered water samples and the fluorometer.

This method performed well overall during both cruises but it performed relatively poorly at one station during each cruise (S187-026 and S199-023). At both stations, the C_{tag} profiles did not match the profiles measured from the water samples. Interestingly, both stations were at approximately the same latitude (7°N) (Table 1), which was at the edge of the equatorial upwelling zone, and both stations had relatively shallow chlorophyll maxima.

Below the euphotic zone, the deep scattering layers influenced the K_{tag} profile. The K_{tag} profiles showed a large secondary peak below the euphotic zone even though there is only a negligible amount of chlorophyll at those depths (Fig. 5). These deeper peaks corresponded well to the depths at which the ADCP echo intensity increased (Fig. 5). The increased echo intensity indicates the presence of deep scattering layers and increased biomass at those depths (Flagg & Smith 1989, Heywood et al. 1991). Based on the ADCP, the deep scattering layers during the oceanographic sta-

tions were at depths ranging from approximately 290 to 440 m. The overall rms difference between the deep scattering layer depths estimated by the tags and the ADCP was 17 m.

Pacific bluefin tuna experiments

For the Pacific bluefin tuna experiments, tag-estimated chlorophyll profiles were comparable to chlorophyll profiles recorded in the WOD for the region. As an example, the irradiance profile collected by bluefin tuna C0071 from 2 June 2005 was used to estimate the K_{tag} and C_{tag} profiles (Fig. 6A). The C_{tag} profile was comparable to the WOD chlorophyll profiles in June that were within 1° of the fish's geolocation (Fig. 6B).

Overall, chlorophyll profiles from the bluefin tuna had maximal chlorophyll concentrations and chlorophyll maxima depths that were similar to those recorded in the WOD for the Baja California region. The rms differences between the C_{tag} at the chlorophyll maxima and the WOD chlorophyll profiles were 0.23 and 0.26 mg m^{-3} for tags sampling at every 4 and 8 s, respectively. Most of the tag-estimated maximal chlorophyll concentrations (91.8% for 4 s tags, 88.5% for 8 s tags) were also within the 95% confidence intervals of the WOD maximal chlorophyll concentrations. In addition, the rms differences in the chlorophyll maxima depths estimated by the tags and in the WOD were 25 and 26 m for the tags sampling at 4 and 8 s, respectively. The majority of tag-estimated chlorophyll maxima depths (82.4% for 4 s tags, 73.0% for 8 s tags) were also within the 95% confidence intervals of the WOD chlorophyll maxima depths.

Daily chlorophyll profiles were concatenated to illustrate the changes in their chlorophyll profiles (see Fig. 6C for an example for bluefin tuna C0071). For instance, the mean chlorophyll concentration at the chlorophyll maxima for bluefin tuna C0071 was $0.45 \pm 0.08 \text{ mg m}^{-3}$ and the mean chlorophyll maxima depth was $69 \pm 14 \text{ m}$. Changes in the C_{tag} profiles were compared to the corresponding horizontal movements, ambient temperature profiles and depth preferences (Fig. 7); however, a detailed analysis of this is beyond the scope of this paper. Depths of the chlorophyll maxima from bluefin tuna C0071 were significantly correlated with the thermocline depth ($R^2 = 0.60$, $p < 0.0001$). The correlation between the thermocline and the chlorophyll maxima depths adds to our confidence in the C_{tag} profiles because the nutricline is closely related to the thermocline in the region and the chlorophyll profile is likely to be related to the nutricline depth (Wilson & Coles 2005).

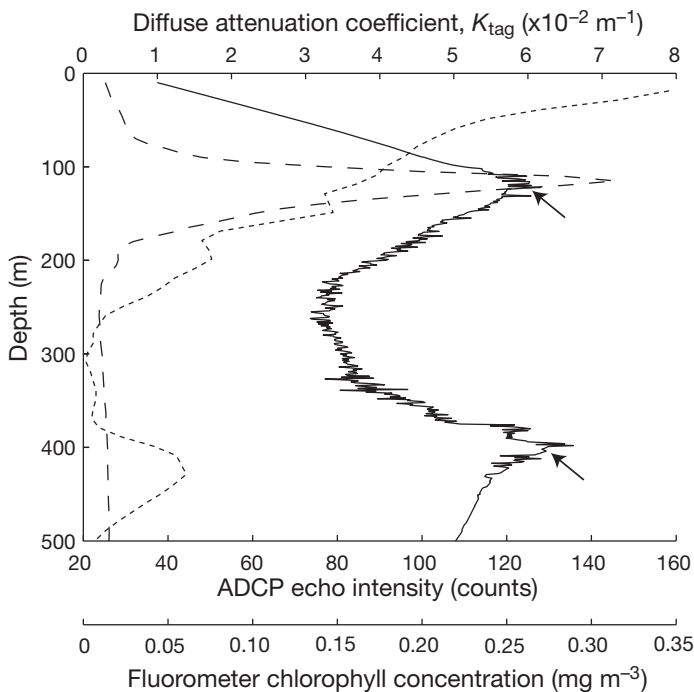


Fig. 5. Profiles of diffuse attenuation coefficients, K_{tag} , estimated from tag C0300 (—), chlorophyll concentrations estimated by the fluorometer (---) and Acoustic Doppler Current Profiler (ADCP) echo intensity (.....) at Stn S199-010. Arrows indicate peaks of K_{tag} at this station, with the shallow peak coinciding with the chlorophyll maxima and the deep peak coinciding with a peak in ADCP echo intensity

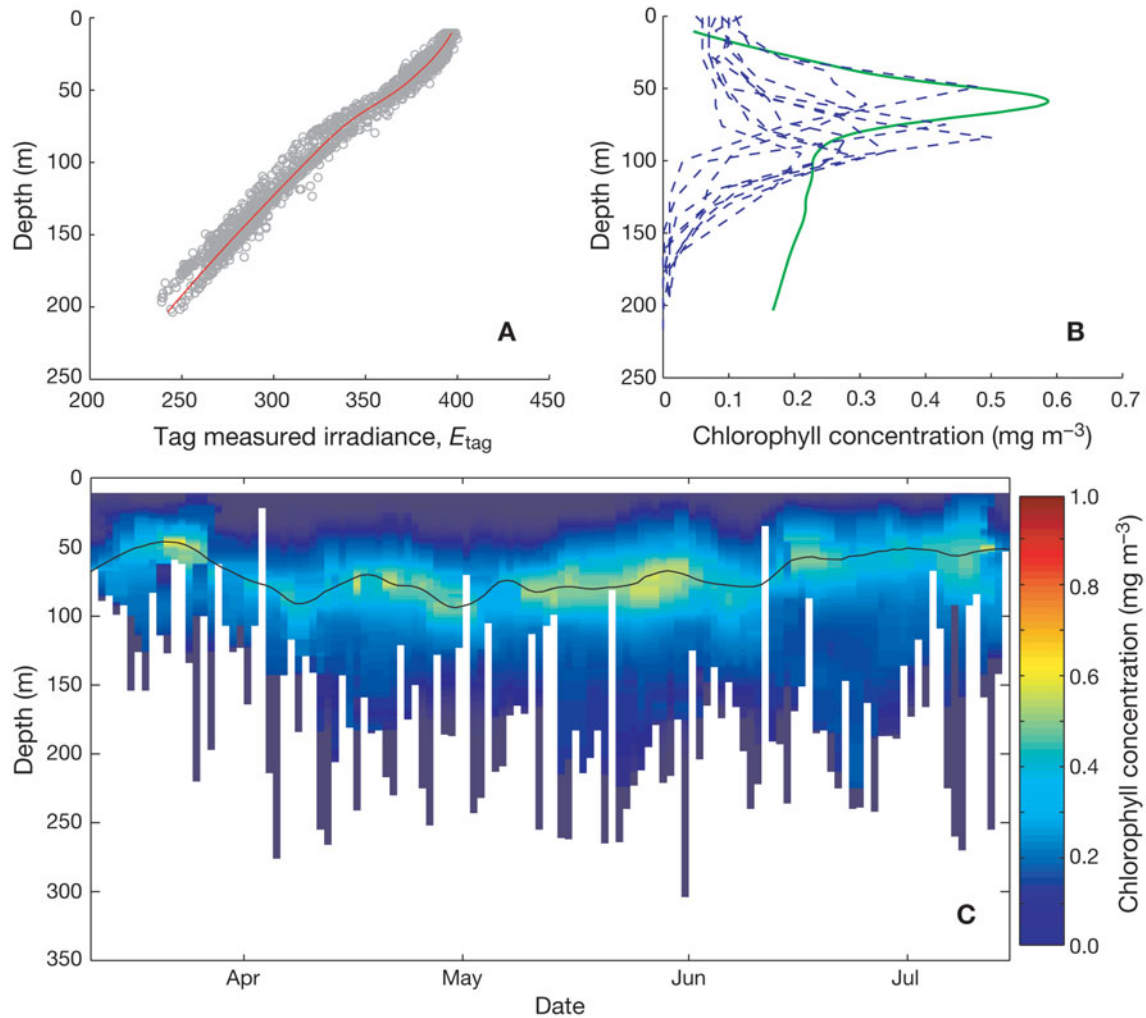


Fig. 6. Irradiance, ambient temperature and chlorophyll profiles estimated from Pacific bluefin tuna C0071 on 2 June 2005 off Baja California. (A) Irradiance (E_{tag} , grey circles) and interpolated irradiance profile calculated using a loess fit (red line). (B) Chlorophyll profiles estimated from the tag (green) and in the World Ocean Database for the month of June and within 1° of the fish's estimated geolocation for 2 June 2005 (dashed lines). (C) Chlorophyll profiles of Pacific bluefin tuna C0071 from 11 March to 15 July 2005. Black line indicates depth of chlorophyll maxima

DISCUSSION

In the present study, we have demonstrated the concept of using light level and depth data from geolocating electronic tags deployed on pelagic animals to obtain *in situ* chlorophyll profiles. Although there clearly remains much more work to be done to improve the accuracy and precision of the chlorophyll estimates, our results suggest that the approach is promising. With further refinements to this approach, research programs utilizing electronic tags can maximize the use of oceanographic data collected by these tags. One potential use will be to obtain subsurface chlorophyll profiles along an animal's path while simultaneously observing the animal's behavior and physiology, which will likely improve our understanding of the

relationship between pelagic animals and the deep chlorophyll maxima. In addition, it also has the potential to greatly increase the number of *in situ* chlorophyll profiles available to the oceanographic community and, hence, improve our understanding of subsurface chlorophyll dynamics.

One of the important sources of error is variability in the ambient light field. If the ambient light field is rapidly changing, any change in the irradiance profile might reflect ambient light field changes rather than light attenuation. To reduce the influence of ambient light field variability, we only used data from a 2 h period around the local noon, which improves the accuracy and reduces the variability of the K_{tag} and, hence, the C_{tag} profiles. Since the rate of change of solar irradiance is at its slowest during local noon, this

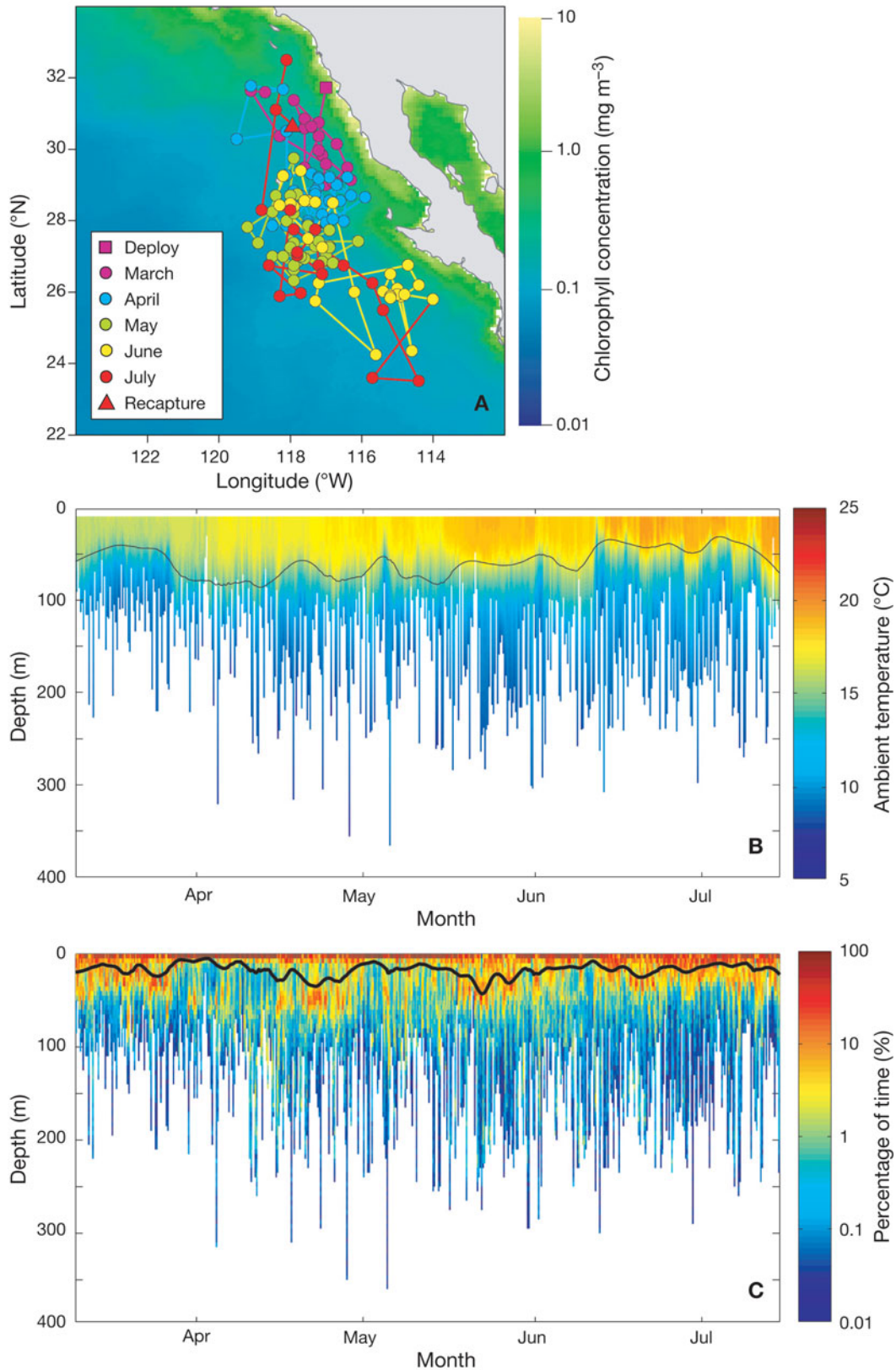


Fig. 7. (A) Horizontal movements. Monthly horizontal movements are shown with mean chlorophyll concentrations for 2005; (B) ambient temperature profiles (black line indicates thermocline depth); and (C) depth preferences of Pacific bluefin tuna C0071 from 11 March to 15 July 2005 (black line indicates mean depth)

would reduce ambient light field changes and hence improve the K_{tag} and C_{tag} profiles. The high solar elevation angle during local noon also reduces the effect of changing atmospheric conditions because the apparent atmospheric optical thickness is reduced. In addition, the high solar elevation angle reduces the reflectance and diffraction of light at the air–sea interface due to Snell's law and the irradiance profile in the water is less influenced by a changing sea state (Kirk 1994). Another advantage is that the solar irradiance, for a given day and set of conditions, is maximal during local noon, which would increase the penetration depth of light and hence the depth range for which this method is effective.

Diving behavior of the animal also interacts with the tag's optical geometry to affect our ability to estimate chlorophyll profiles. Here, we used the LTD2310 tag, which has a flexible cylindrical light collector that protrudes into the water column and moves around rapidly as the animal swims. Hence, the tag can be assumed to measure overall scalar irradiance (i.e. measuring light equally from all directions). Therefore, the angle at which the photons intercept the light sensor does not affect the tag's irradiance measurement. On the other hand, cosine collectors (flat light sensors that face upwards or downwards) are sensitive to the angle at which the photons intercept the light collector. Therefore, tags with cosine collectors are more likely to be sensitive to changes in the tag's orientation during an animal's dive. If the orientation of the animal during a dive is relatively random, changes in the optical geometry should equal out over time. However, if the animal's orientation changes in a regular fashion during specific parts of the dive, it would be important to correct the irradiance data for body orientation.

For this method to be effective when the pelagic animals are moving rapidly over large spatial scales at relatively short temporal scales, it will likely require modifications to the algorithm described here. When pelagic animals are moving rapidly, the animal may be moving through several water masses within a short time period. Therefore, the light attenuation profile data from a single day may consist of a mixture of light attenuation profiles from different water masses. One possible way to tackle this problem would be to focus the analysis on an individual dive over a short time period. In this way, the chance of the data being from multiple water masses is reduced. The drawback of this approach, however, is that the tag would need to be sampling at very high rates for there to be enough light and depth data for analysis.

The relatively poor performance of this method at the oceanographic stations close to the upwelling zone suggests the need for potential improvements to the current method. In addition to increased primary pro-

ductivity, upwelling zones also tend to have increased secondary productivity, especially near the edges of the upwelling zones (Cury & Roy 1989). The increased zooplankton abundance in upwelling areas may have led to a reduction in the performance of Morel's (1988) bio-optical model, which was parameterized for Case-I (open ocean) waters. One way to improve the future performance of this method is to measure the irradiance and attenuation at multiple optical bands, rather than a single optical band. For example, the ratio between the remote sensing reflectance at 490 and 555 nm is used in empirical and semianalytic bio-optical models to estimate the surface chlorophyll content from remote sensing data (O'Reilly et al. 1998). Since the remote sensing reflectance at a specific wavelength is related to K_d (Mobley 1994), we will be able to use a similar multi-band approach to estimate chlorophyll profiles from electronic tags. Indeed, we are currently testing the use of dual-band tags on northern elephant seals to estimate chlorophyll profiles. These multi-band instruments will also allow us to use new bio-optical algorithms developed for passive optical sensors deployed on AUVs and Argo floats (Brown et al. 2004).

Besides Morel (1988), there are alternative bio-optical models that can be used with electronic tags and it would be useful to compare the performance of these different models in future studies. However, it would be important that the details of each model be understood and found to be applicable for the study. For example, Morel & Maritorena (2001) used new optical data from several Joint Global Ocean Flux Study (JGOFS) cruises to revise Morel's (1988) bio-optical model. However, we were unable to use those results in the present study because the *in situ* chlorophyll concentrations from the JGOFS cruises were primarily measured using HPLC (Morel & Maritorena 2001). In contrast, the chlorophyll concentration data from Morel (1988) were primarily determined using fluorometric methods, which were similar to the methods used in the present study. In addition, it would be useful to compare the performance of electronic tags from different manufacturers under different environmental conditions. Since the optics and electronics of tags from different manufacturers can differ substantially, we expect that this may result in differing abilities to estimate chlorophyll concentrations.

Bio-optical models are typically developed for surface waters, but we showed that K_{tag} profiles below the euphotic zone might be used to estimate deep scattering layer depths. The K_{tag} profiles often showed a large secondary peak below the euphotic zone even though there was only a minimal amount of chlorophyll at those depths. These deeper peaks corresponded to the depths at which ADCP echo intensity increased

(Fig. 5), suggesting that scattering and absorbance of light by animals in the deep scattering layers were detected by the tags. However, caution should be used in interpreting this result because the irradiance levels at these depths were approaching the detection limit of this generation of tags. We speculate that the next generation of tags, with potentially 2 to 3 more orders of magnitude of sensitivity to light, might be used to validate the ability to estimate deep scattering layer depths. In addition, it would be important to test this by deploying tags on species that are associated with the deep scattering layer and dive repeatedly through the water column during the day, e.g. bigeye tuna and elephant seals (Dagorn et al. 2000, Le Boeuf et al. 2000, Campagna et al. 2001, Schaefer & Fuller 2002). In the future, it would also be interesting to correlate the biomass at these depths with the K_{tag} profiles obtained from the tags. This would enable us to estimate the deep scattering layer biomass with these tags and allow us to determine if long-term changes in diving behavior are related to changes in the deep scattering layer depth and biomass. Currently, we have to deploy ships or AUVs to obtain acoustic profiles or net samples of the deep scattering layer at the same time as the animal is being tracked (Dagorn et al. 2000).

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